Estimation of real evapotranspiration and its variation in Mediterranean landscapes of central-southern Chile

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\textbf{A B S T R A C T}

Evapotranspiration (ET\textsubscript{d}) is a key controller in the ecohydrological processes of semi-arid landscapes. This is the case in the dry land in Chile’s central-southern zone, where forestry, farming and livestock activities must adapt to precipitation with considerable year-on-year variations. In this study, the spatial distribution of ET\textsubscript{d} was estimated in relation to the land use maps and physical parameters of the soil. The ET\textsubscript{d} was estimated through the Simplified Surface Energy Balance Index (S-SEBI) using data from weather stations and remote data provided by the ASTER and MODIS sensors for November 2004 and 2006, respectively. The spatial variability of ET\textsubscript{d} was compared among different plant types, soil textural classes and depths using non-parametric statistical tests. In this comparison, the highest rates of ET\textsubscript{d} were obtained in the forest covers with values of 7.3 ± 0.8 and 8.4 ± 0.8 mm d\textsuperscript{-1} for 2004 and 2006, respectively. The lowest values were estimated for pastures and shrublands with values of 3.5 ± 1.2 mm d\textsuperscript{-1} and for crops with rates of 4.4 ± 1.6 mm d\textsuperscript{-1}. Comparison of the ET\textsubscript{d} of the native forest covers and plantations of exotic species showed statistically significant differences; however, no great variation was noted, at least in the study months. Additionally, the highest rates of ET\textsubscript{d} were found in the clay loam textures (6.0 ± 1.8 and 6.4 ± 2.0 mm d\textsuperscript{-1}) and the lowest rates in the sandy loam soils (3.7 ± 1.6 and 3.9 ± 1.6 mm d\textsuperscript{-1}) for 2004 and 2006, respectively. The results enable analysis of the spatial patterns of the landscape in terms of the relation between water consumption, ET and the biophysical characteristics of a Mediterranean ecosystem. These results form part of the creation of tools useful in the optimization of decision-making for the management and planning of water resources and soil use in territories with few measuring instruments.

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1. Introduction

Evapotranspiration (ET\textsubscript{d}) is a key component in the processes that guide the interactions of water and energy in semi-arid ecosystems. The water balance in these ecosystems is regulated by solar radiation and pluviometric events characterized by their low frequency and varied intensity, causing moisture in the soil that restricts soil-plant-atmosphere interactions. Under these conditions, ET\textsubscript{d} represents a large proportion of the water budget, estimated at more than 70% of incoming precipitation (Kurc and Small, 2004; Huxman et al., 2005; Breshears, 2006; Moussa et al., 2007). In such landscapes, the soils store the incoming water from precipitation for potential biological activity (Huxman et al., 2004). This means that the variation in the availability of water resources is closely linked to changes in the physiological and structural states of the vegetation, which may be important in the control of ecological processes. Accordingly, ET\textsubscript{d} is the central controller in ecohydrological processes in semi-arid environments, such as productivity of the ecosystem (Huxman et al., 2005; Yepez et al., 2005) and the influences of the vegetation on the water and the energy exchange (Moreira et al., 1996). Despite advances in the understanding of these phenomena, it has not yet been fully documented how the surface and subsurface characteristics of the soil can modify water and energy flows.

The Mediterranean landscape in Chile’s central-southern zone is characterized by vast expanses of forest plantations and rainfed agricultural crops, mainly wheat, that must adapt to the rain cycles, which are scarce, intermittent and with significant year-on-year variations, as is frequently the case in semi-arid contexts (Newman et al., 2006). In addition to the water limitations during the dry season due to the absence of precipitation, water availability problems have intensified in recent decades. This is due to an increase in the demand for water for agricultural uses and human consumption (Lara et al., 2003), which may be critical in future
scenarios given the climate change trends that project a decrease in precipitation and an increase in temperature in these territories (Fuenzalida et al., 2006).

Moreover, this zone has undergone strong man-made modifications, a product of the expansion of plantations of exotic forest species (mainly *Pinus radiata*), which has led to the elimination of extensive areas of native forest. This decrease has been estimated at 67% of the initial surface of native forest between 1975 and 2000 (Echeverría et al., 2006). These modifications in soil use and forest cover are important to consider in the hydrological cycle, and these have been studied regarding their effect on ET$_d$ and water consumption (Bosch and Hewlett, 1982; Iroumé and Huber, 2002; Iroumé et al., 2006; Huber et al., 2008, 2010; Birkinshaw et al., 2011). Huber et al. (2008) reported greater water consumption (ET$_d$) on plantations of exotic species (*P. radiata* and Eucalyptus spp.) than on pasture and shrublands. These works focused mainly on catchment basins of less than 100 ha due to the greater control in the samplings, which increases their likelihood of success given that areas with fragmented landscape and reliefs could bias the results (Bosch and Hewlett, 1982). By contrast, in catchment basins of 100–1000 km$^2$, Pizarro et al. (2006) and Little et al. (2009) documented a decrease in the production of water flows due to the growth of forest plantation covers of exotic species to the detriment of the native forest.

Therefore, it is relevant to conduct studies on the spatial distribution patterns of ET$_d$ and its relationship with various land use maps. Therefore, determining how ET$_d$ is related to this information contributes to the understanding of processes that explain landscape hydrology, particularly on vast expanses and at a territorial level in the central-southern zone of Chile.

Surface energy balance models based on satellite images have successfully been applied to estimate the spatial distribution of ET$_d$ in various landscapes. The Simplified–Surface Energy Balance Index (S-SEBI) (Roerink et al., 2000) is an algorithm that can estimate energy flows and ET$_d$ through spatial contrasts between captured hydrological conditions and the information from satellite images regarding surface reflectance and the thermal region of the spectrum. This model requires a minimum of meteorological data, adapting to zones with few measurements in situ or that simply do not have these measurements. The S-SEBI model has been widely applied and evaluated successfully in obtaining the ET$_d$ with different satellite sensors on a wide variety of ecosystems and at different spatial scales (Roerink et al., 2000; Gomez et al., 2005; Sobrino et al., 2005, 2007, 2008; Verstraeten et al., 2005; García et al., 2007, 2008; Boronica and Ramíllien, 2008; Galleguillos et al., 2011; Mattar et al., 2013) on flat as well as mountainous terrains.

The spatial estimation of ET$_d$ has been of interest to evaluate the characteristics of the landscape serving as an indicator of the water deficit of the surface (García et al., 2007, 2008), in the classification of functional types of ecosystems (Fernández et al., 2010), in the modeling of soil attributes (Taylor et al., 2013), and in the characterization of types of soil cover (Péças et al., 2013).

Considering the aspects analyzed previously and the biophysical variables of the landscape that affect water consumption, the main objectives of this study focus on the dry landscape of the central-southern zone of Chile, (i) to evaluate the spatial distribution of ET$_d$ in relation to the different land use maps and physical variables of the soil that determine the water storage capacity available for plants and (ii) to analyze the relevance of using ET$_d$ to improve the understanding of spatial patterns throughout the territory.

The manuscript is structured as follows: Section 2 describes the study area, the data used and the methodology of this work. Section 3 presents the results and discussions on obtaining the spatial distribution of ET$_d$ and its comparison with the biophysical variables associated with the vegetation and the physical properties of the soil. Finally, Section 4 presents the conclusions.

## 2. Data and methodology

### 2.1. Study area

The study area is two zones of the semi-arid sector of the Maule Region in central-southern Chile, located between 35°00′ and 35°50′S, which cover a total area of 6050 km$^2$. This landscape has been drastically modified in recent decades due to a great expansion of plantations of exotic forest species (mainly *P. radiata*), which led to the elimination of extensive areas of native forests (San Martín and Donoso, 1997), where Echeverría et al. (2006) estimated an annual reduction in native forest at a rate of 4.5%. The heavy development of forest activity has relegated the native forests to small patches (<100 ha) of secondary forests comprised mainly of *Nothofagus* (*N. obliqua* and *N. glauca*) and sclerophyll species such as *Accacia caven*, *Quillaja saponaria* and *Maytenus boaria* (Echeverría et al., 2006).

The area presents a Mediterranean climate with winter rains and a prolonged dry season between 7 and 8 months. Annual precipitation varies between 600 and 900 mm per year, concentrated mainly between May and August. The potential evaporation during the summer months is between 200 and 500 mm, exceeding the scarce rain. This zone is described by Little et al. (2009) as a semi-arid zone; however, according to the criteria of De Pauw et al. (2000) it is a sub-humid zone. In spite of this, the climatic conditions make the zone subject to a significant water deficit, which is critical in the dry season, with low cloud cover and high luminosity. These conditions enable the development of rainfall crops in the study area, mainly wheat, which must adapt to the scarce, intermittent rain cycles that vary considerably year-on-year.

To ascertain the topography of the terrain in the study area, a digital elevation model (DEM) was used via the ASTER GDEM, which has a spatial resolution of 30 m. This product was developed jointly by NASA and the Japanese Ministry of Economy, Trade and Industry (METI) and is freely available at http://www.gdem.aster.ersdac.or.jp/. The DEM was used to generate the slope and exposure maps needed to estimate solar radiation. In Fig. 1, an infrared composition appears on the DEM in the study area.

### 2.2. Spatial remote data and GIS

Two ASTER Level-2 product scenes were used of surface reflectance (AST-07), bands of surface emissivity (AST-5) and surface temperature (AST-08), which already have radiometric and atmospheric corrections (Abrams, 2000). The images of surface reflectance corresponding to the bands in the visible (Vis) and near infrared (NIR) were obtained at 15 m pixels, the bands in the shortwave infrared (SWIR) were obtained at 30 m pixels, while the products AST-5 and AST-8 corresponding to the bands in the thermal infrared (TIR) were obtained at 90 m pixels. The two scenes were acquired for November 5 2004 and November 18 2006 around 3 p.m. UTC (12 p.m. local time) under clear skies. The two scenes correspond to the period of maximum vegetation growth, where there are conditions of high radiation as well as optimal conditions in terms of water availability in the soil, this being immediately after the months where most of the rain is concentrated. Indeed, it has been shown in studies on Mediterranean ecosystems (Galleguillos et al., 2011) that during the spring-summer months the spatial differences of ET$_d$ are maximized when there is a water deficit of varying magnitudes in the vegetation caused by the progressive exhaustion of the water reserves in soils. This phenomenon has greater impact on soils with physical properties that involve a lower retention capacity and subsequent delivery of water to the plants.

For the ASTER scenes, the Normalized Difference Vegetation Index (Rouse et al., 1973; Tucker, 1979) was calculated using the
The air temperature at surface level was obtained by interpolating the atmospheric profile (MOD07) in the first 1500 m of elevation (levels of pressure greater than 850 hPa) into the level of surface pressure, assuming a hydrostatic atmosphere, according to the equation:

\[ T_a = \frac{P_i - P_{surf}}{P_i - 850} T_{850} + T_i \]  

(1)

where \( P_i \) is the level of pressure closest to the surface where data can be obtained (1000, 950 or 920 hPa), \( P_{surf} \) is the surface pressure, \( T_i \) is the air temperature at the level of atmospheric pressure \( i \), and \( T_{850} \) is the temperature at the 850 hPa level. The product obtained from the air temperature was resampled to a size of 90 m × 90 m pixels, using a regridding method based on cubic convolution, which consists of a non-linear interpolation as suggested by Zhao et al. (2005) for meteorological data.

The global solar radiation was obtained on the basis of the Bristow and Campbell (1984) model and described in detail in (Olivera et al., 2013). Although this model was defined for daily solar radiation values, it was evaluated successfully for monthly mean values (Meza and Varas, 2000), which is why this work used a new set of coefficients for an instantaneous time scale, calibrating the model for hourly data with the measurements from 2:30 p.m. and 3 p.m. (UTC) from the Pananillos station for the clear days in November of 2010 and 2011. The model obtained a \( R^2 \) coefficient of 0.69 and a root mean square error (RMSE) of 48.7 W m\(^{-2}\) (5.1\% relative error) in the calibration.

2.5. Estimation of daily evapotranspiration

To estimate evapotranspiration, the S-SEBI algorithm was used, which is an approach based on the balance of surface energy, which can be represented as follows:

\[ \lambda ET_i = R_{ni} - G_{li} - H_i \]  

(2)

where \( R_{ni} \) is the net radiation resulting from the incident and outgoing shortwave and longwave radiation [W m\(^{-2}\)]; \( H \) is the sensible heat flow [W m\(^{-2}\)]; \( G \) is the heat flow of the soil [W m\(^{-2}\)]; \( \lambda ET \) is the latent heat flow [W m\(^{-2}\)]; \( \lambda \) is the latent heat from vaporization of the water, approximately 2450 J g\(^{-1}\) at 20 °C; and \( ET \) represents the evapotranspiration [g m\(^{-2}\) s\(^{-1}\)]. The subindex ‘i’ refers to the
instantaneous values whereas the subindex ‘d’ is used for daily values.

To solve the components of Eq. (2), the S-SEBI model was used (Roerink et al., 2000) based on the surface contrasts that give account of the variability in hydrological conditions within the study area. These contrasts allow the theoretical limits, wet (TH) and dry (Te), to be determined from a scatter plot of surface temperature values (Ts) and surface albedo (αs) in order to derive the instantaneous ETd pixel by pixel from the evaporative fraction (EF).

This approach can be expanded to estimate daily evapotranspiration ETd, which assumes EFd ≈ EFd and the negligible daily heat flow of the soil (Cgd ≈ 0), whereas the net radiation at the moment the satellite passes (RN) can be extrapolated to a daily scale using the ratio Cgd/RN, which depends on the hour and day of the year (Sobrino et al., 2007). So ETd can be expressed according to the following expression:

\[
ET_d = 86400 \cdot EF_a \frac{RN}{\lambda} = 86400 \cdot \frac{T_H - T_s}{T_H - T_{eT}} \cdot \frac{C_aR_m}{\lambda}
\]

The net radiation was derived from the ASTER and MODIS data using the general equation (Verstraeten et al., 2005; Chehbouni et al., 2008):

\[
RN = (1 - α_s)R_g + ε_s(σR_a T_{a}^4 - σT_s^4)
\]

where αs is the surface albedo estimated according to Liang (2001) from the linear combination of the reflectance bands of the product AST-07 averaged to 90 m; Rg is the global solar radiation at the moment the satellite passes [W m⁻²]; εs is the surface emissivity calculated from the emissivity bands of the product AST-05 (Ogawa et al., 2003); εa is the atmospheric emissivity [–]; calculated from an empirical formulation in terms of the air temperature Ta (Bastiaanssen et al., 1998); σ is the Stefan–Boltzmann constant (5.67 × 10⁻⁸ [W m⁻² K⁻⁴]), and Ts is the surface temperature obtained directly from the product AST-08 [K].

Table 1
Meteorological data available for each weather station.

<table>
<thead>
<tr>
<th>Weather station</th>
<th>Lat (°)</th>
<th>Long (°)</th>
<th>Altitude (m.a.s.l.)</th>
<th>2004</th>
<th>2005</th>
<th>2006</th>
<th>2010–2011</th>
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<tr>
<td>Lico</td>
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<td>72.07</td>
<td>10</td>
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<tr>
<td>Curico</td>
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<td>71.14</td>
<td>225</td>
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<td></td>
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<tr>
<td>Huquen</td>
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<td>150</td>
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<tr>
<td>Panguilemo</td>
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<td>113</td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Talca</td>
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<td>71.37</td>
<td>100</td>
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<td></td>
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<td>San Pedro</td>
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<td>353</td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Dunas De Chanco</td>
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<td>72.34</td>
<td>35</td>
<td></td>
<td></td>
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<td></td>
</tr>
<tr>
<td>Botacura</td>
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<td>72.13</td>
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<td></td>
<td></td>
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<tr>
<td>Chanco</td>
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<td>72.33</td>
<td>60</td>
<td></td>
<td></td>
<td></td>
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<tr>
<td>Verbas Buenas</td>
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<td>150</td>
<td></td>
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<tr>
<td>Pantanillos</td>
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<td>71.30</td>
<td>312</td>
<td></td>
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<td>***</td>
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<tr>
<td>Nirivilo</td>
<td>35.53</td>
<td>72.08</td>
<td>200</td>
<td></td>
<td></td>
<td></td>
<td>***</td>
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<tr>
<td>Constitución</td>
<td>35.32</td>
<td>72.40</td>
<td>10</td>
<td></td>
<td></td>
<td></td>
<td>***</td>
</tr>
</tbody>
</table>

* Daily minimum and maximum air temperature data (2 observations per day).
** Daily precipitation data for the whole year.
*** Air temperature, solar radiation and relative humidity each 30 min.
2.6. Relationship between landscape and evapotranspiration

Once the $ET_d$ was estimated from the balance of surface energy, it was compared to the biophysical variables of the land use map and physical properties of the soil: textural classes and depth.

The variation of the $ET_d$ according to the variable landscape was analyzed from boxplots, for vegetation, soil texture and depth. For each variable, a sample was constituted from all the pixels that belonged to a class, so that for the ground cover 10 samples were compared, and for the properties of soil texture and depth, 5 samples for each were compared.

The spatial information was pre-processed, based on the elimination of zones where the relations were confused, such as those sectors that underwent strong degradation as a result of deforestation or logging, those that were visualized as bare soils or sectors with young vegetation. This filter made it possible to analyze in detail the $ET_d$ patterns among the different forest covers.

In this comparison the behavior between classes of each variable of the landscape studied was analyzed, examining the significant differences existing between classes with the Kruskal–Wallis test of non-parametric contrasts and later by means of the Mann–Whitney non-parametric test to compare class with class. This study employed a 95% confidence level ($\alpha = 0.05$). To compensate for the Type I error due to the multiple comparisons in the Mann–Whitney contrast, the $\alpha$ level was divided by the number of comparisons according to each variable. It has to be noted that the statistical analysis was realized under the concept of areal model as described by Cressie (1991). This approach was chosen according to the ground data available which correspond to discrete data. Each polygon (or group of polygons) that represents a ground variable, gathering consequently thousands of pixels of $ET_d$, which were averaged for each class, diminishing significantly the probability of auto-correlation between $ET_d$ and the classes.

3. Results and discussion

3.1. Daily evapotranspiration

Once the S-SEBI method had been applied to the two ASTER scenes, it was possible to obtain distribution maps of the $ET_d$ on the dry land of the Maule Region (Fig. 3).

It is observed that in the 2006 scene higher rates were obtained than in that of 2004, so that in the 2004 scene the mean $ET_d$ was 5.82 mm d$^{-1}$ ($\pm 1.76$) whereas in the 2006 scene the mean reached 6.39 mm d$^{-1}$ ($\pm 2.08$). This difference is due to meteorological input data, higher air temperature and global solar radiation, and can also be explained by a greater water availability due to the greater precipitation in 2006. Thus, the precipitation recorded in the period from April to the date of each image at the Constitución station was 751.0 and 929.3 mm for 2004 and 2006, respectively. In the same zone, the Nirivilo station presented values of 700.2 and 844.5 mm for the same years. This is why in 2006 the precipitation increased by 20.6% and 23.1% at both stations compared to 2004 (Fig. 4). This may reflect that there was a greater amount of water available in the soil, at least in the deepest layers, for the deep-rooted vegetation, and this will be discussed in further sections.
3.2. Relationship between daily evapotranspiration and vegetation

In a Fig. 5 box plot comparison is made for each scene between classes of the land use map, where practically the same trends can be observed between classes for both scenes.

In both scenes the highest $ET_d$ rates (Figs. 3 and 4) were found in forest cover (Class 7–9), with means between 7.30 and 8.39 mm d$^{-1}$. This can best be seen in the comparison by box plot for the two scenes, in which the highest means appear (7.30–7.46 mm d$^{-1}$ for the 2004 scene and 8.01–8.39 mm d$^{-1}$ for the 2006 scene). Despite having very similar means, the classes show $ET_d$ rates with statistically significant differences in both scenes, according to the Mann–Whitney non-parametric test, with the rates of forest plantations being higher than the native forests, and these in turn being higher than the mixed forests (Class 7 > Class 8 > Class 9). Similar studies have already shown these differences, where a greater water consumption by forest plantations, mainly of $P$. radiata, than by native forests has been determined (Huber et al., 2008; Lara et al., 2009; Little et al., 2009). In spite of the similarity in the results, in previous works these were obtained by estimating the water balance and water flow production. In this work equivalent results were established using remote sensing and spatially distributed energy balance estimation models.

The lowest rates of $ET_d$ can be observed near the coast, in the southwestern zone of the 2004 scene and the northwestern zone of the 2006 scene, where there is terrain without vegetation (dunes), and in the interior of each scene (right), where rainfed agricultural activities, pastures and shrublands with less vegetation cover are developed. Fig. 4 illustrates the effect of vegetation on the amounts of $ET_d$ such that the agricultural lands (Classes 1 and 2) present rates much lower than those obtained in forests, presenting similar means to meadowland, between 4.37 and 4.95 mm d$^{-1}$ (Table 2). These low amounts of evapotranspiration in the crops correspond mainly to rainfed wheat. These crops are frequently found under conditions of water stress, mainly in the phenological stages after tasseling, corresponding to the dates of the satellite scenes where the plants are beginning the final maturation process with the resulting drying out of their plant structures. This significantly increases the sensible heat flow (and therefore the latent heat or $ET_d$ decreases) when there is a low rate of structural humidity and physiological activity linked to canopy transpiration.

In terms of shrub-pasture, shrubland and arboreous shrubland covers (Classes 4–6), these show an increase in $ET_d$ due to the increase in shrub and tree cover. This may be due to the water contained in the upper layers of soil beginning to dry up on these dates, which is why shrubs and trees have more water available in the deeper layers. Zhang et al. (2001) and Huxman et al. (2005) have determined that the invasion of ligneous plants into ecosystems changes the components of the hydrological cycle, reducing the flows generated and increasing the transpiration of the vegetation when using the water available in the soil. At the same time, they suggest that these changes are more relevant in semi-arid ecosystems, being small or negligible in arid and sub-humid environments.
environments. This is reflected in the study area, where the forest plantations present the greatest evapotranspiration rate compared to the rest of the surface covers. With different methods and for the same study area, similar results were reported by (Huber et al., 2008), where similar differences were found between plantations of *P. radiata* compared with pastures and shrubs, also demonstrating a reduced percolation and high water consumption due to evapotranspiration.

To determine the influence of vegetation on ET$_d$ rates quantitatively, the NDVI spectral vegetation index was compared with the ET$_d$ through a scatter diagram (not shown), obtaining a linear regression with a high correlation ($R^2$ of 0.76 and 0.78 for the 2004 and 2006 scenes, respectively). Therefore, the NDVI might explain in large measure the ET$_d$ recorded in the zone, as has been demonstrated in works on various ecosystems (Yang et al., 2006; Cleugh et al., 2007; Mu et al., 2007; Wang et al., 2007; Glenn et al., 2008). Additionally, the NDVI is an easy variable to obtain through remote sensing, having the advantage that surface reflectance is a greater spatial resolution than thermal images required in the estimation of energy flows.

With respect to the areas without vegetation (Class 10), the great variability recorded stands out, where the quartiles are distributed practically between the minimum and maximum amounts of ET$_d$ found in each scene. This situation is explained basically by the presence of vegetation, bodies of water and wetlands formed in the middle of the dunes in both scenes, classified according to the GIS as cover within Class 10. This occurs because the minimum mapping area of 6.25 has, where dunes coexist, is composed of dry zones with minimum ET$_d$ rates, both as wetlands and bodies of water, the ET$_d$ rates of which are the maximum.

According to the test of contrasts between classes, the Mann–Whitney test revealed that the only classes that did not show any significant differences in their measures of central tendency were the crop-pasture rotation and pasture, croplands and shrubland-pasture in the 2006 scene. These results pointed out the logical response in water consumption between vegetation that has ecological similarities in terms of structural and functional properties such as architecture, height of individual, depth rooting and fenological periods that depends mainly on rainy season. In this case, the last precipitation event was registered 36 days before the satellite overpass (Fig. 4), even though that in 2006 the total precipitations were higher than in 2004. Therefore the soil surface was probably dried enough to undifferentiated the effect of water consumption, expressed as ET$_d$, between all types of vegetation classes that interact with available water stored in the soil. It may represent one of the uncertainties of the proposed approach.

Meanwhile, in the 2004 scene all the results between classes showed statistically significant differences. These results highlight the relevance of ET$_d$ as an ecosystem functional tool (García et al., 2008; Fernández et al., 2010), able to discriminate among all the land use classes. In this condition, where a 28 mm rain was registered 11 days before the ASTER scene, soil moisture was enough to allow transpiration of non-stressed plants, highlighting the physiological and structural capacities of the vegetation to produce different ET$_d$ rates.

### 3.3. Relationship between daily evapotranspiration and physical properties of the soil

Figs. 6 and 7 illustrate the comparison using boxplots for each scene between textural classes and soil depth, respectively. In terms of the ET$_d$ patterns according to textural classes, with the exception of the sandy textures, confirms the theoretical trend of the relationship between textures and the water exchange capacity, recording very similar patterns between classes in both scenes. The highest ET$_d$ rates were found in the clay loam textures with means of 6.04 (±1.76) and 6.40 (±1.97) mm d$^{-1}$ and the lowest rates in the sandy loam soils with means of 3.72 (±1.57) and 3.95 (±1.59) mm d$^{-1}$ for the 2004 and 2006 scenes, respectively.

This tendency between classes is explained basically by the different capacity that the soils have in delivering available water to the vegetation and which is conditioned in part by this physical attribute. Thus, the finest textures presented greater surface per unit of volume, enabling a greater adsorption of water films by soil particles. Nevertheless, it is precisely due to this phenomenon that the water is heavily retained in the soil, limiting absorption by the roots, which leads to the resulting decrease in the plant transpiration rate. It is for this reason that loam textural classes are those with the greatest water availability for the plants, since they present the greatest amount of micro and macropores, and the coarse textural classes (sand and sandy loam) present the lowest quantity of available water. However, the sandy class did not exhibit the expected tendency, registering elevated ET$_d$ rates, which might be associated with factor related to vegetation, bodies of water and small wetlands and meadows that can be found in the middle of the sand dunes, as previously mentioned for the cover of areas without vegetation. A comparison of the coarse classes in both scenes revealed that the 2004 scene presented higher ET$_d$ rates, since attempts have been made to control the movements of sand through the extensive plantation of marram grass (*Ammophila arenaria*) (Castro, 1998), as well as the construction of the Federico Albert National Reserve. This reserve contains forest species (*Eucalyptus globulus, P. radiata, Cupressus macrocarpa, Acacia melanoxylon*) introduced to control the expansion of the dunes on the town of Chanco (Paskoff and Manriquez, 1999). This is why this
textural class presents a greater $ET_d$ due to the NDVI increase over the sand cover.

Similar tendencies were found in both scenes (2004 and 2006) in the $ET_d$ by textural class. Previous studies have focused on determining the effect of soil textures on the water properties of those soils, defining the relationship between soil moisture, load and hydraulic conductivity (Fernandez-Illescas et al., 2001; Gutmann and Small, 2005; El Maayar and Chen, 2006). Thus, this movement controls the water balance between evapotranspiration and runoff. Despite these relationships between texture and $ET_d$, Gutmann and Small (2005) showed that the textural classes alone explain between 4% and 14% of the energy flows used in $ET$.

In box plots by soil depth class (Fig. 7), the $ET_d$ did not present the same patterns between classes in the two scenes; the greatest differences were found between scenes in the thinnest soils. Nevertheless, in both scenes it was observed that the thinnest soils (<75 cm) exhibited the lowest rates of $ET_d$, the minimum being in the soils between 25 and 50 cm with rates of 3.84 (±1.24) and 4.13 (±1.53) mm d$^{-1}$ for the 2004 and 2006 scenes, respectively. The soils between 75 and 100 cm had the highest evapotranspiration rates, with means of 6.45 (±1.55) and 7.11 (±1.77) mm d$^{-1}$ for the 2004 and 2006 scenes. This is due mainly to the greater water storage capacity of deep soils due to the greater volume per ground surface. The deepest soils (>100 cm), however, did not present the highest rates, being exceeded by those previously mentioned. This can be explained by the type of vegetation developed in each depth class, since, as can be seen in Table 3, the soils between 75 and 100 cm are covered mainly by forests (996.5 and 651.7 km$^2$) whereas >100 cm have a large proportion of surface covered by agricultural crops and pastures and shrublands (456.2 and 638.6 km$^2$). These deeper soils presented vegetation with short phenological cycles, strongly dependent on surface water; therefore, for the

![Fig. 6. Box plot of $ET_d$ per class of texture of the soil for each scene.](image)

![Fig. 7. Box plot of $ET_d$ per class of soil depth for each scene.](image)
study dates the transpiration rate was much lower than in sectors with forests and soils that were not as deep. These results are consistent with what was reported by Echeverria et al. (2007), who compared the water consumption of a pasture and a native forest in southern Chile. In this study the marked seasonal fluctuation of the moisture content in the soil was verified, with the pasture being the one that presented the most intense drying patterns in the first 100 cm of depth.

The two scenes had dissimilar trends for the \( E_{T_d} \) based on the soil depth as illustrated in Fig. 7. Here it can be seen that for the 2006 scene, the three thinnest soil classes (classes < 75 cm) did not present significant differences in their rates, as were found in the 2004 scene. This can be explained in part by the type of vegetation developed and the pluviometric events recorded before each date of analysis. This way, for the 2006 analysis date no significant precipitation was recorded on the previous days (3.5 mm) (Fig. 4), with the last relevant rain being one month prior to the acquisition date. This resulted in significant drying of thin soils (classes < 75 cm), causing low evapotranspiration rates at the time of the scene. By contrast, for the 2004 scene, significant differences were found in the \( E_{T_d} \) between all the classes, including those of surface soils, when conditioned by the significant precipitation (29.6 mm) recorded ten days before the acquisition of the scene (Fig. 4). This may indicate that there was sufficient water available in the thinnest soils, which is why the differences between classes might be due to the water consumption by the type of vegetation present.

Table 3

<table>
<thead>
<tr>
<th>Land cover type</th>
<th>Date</th>
<th>Depth class [cm]</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td>&gt;100</td>
</tr>
<tr>
<td>Agricultural land (AL, CPR)</td>
<td>2004</td>
<td>238.7</td>
</tr>
<tr>
<td></td>
<td>2006</td>
<td>225.4</td>
</tr>
<tr>
<td>Pastures and shrubland (Pas, SP, Shr, AS)</td>
<td>2004</td>
<td>456.2</td>
</tr>
<tr>
<td></td>
<td>2006</td>
<td>638.6</td>
</tr>
<tr>
<td>Forests (ESP, NF, MF)</td>
<td>2004</td>
<td>831.6</td>
</tr>
<tr>
<td></td>
<td>2006</td>
<td>613.5</td>
</tr>
<tr>
<td>Total</td>
<td>2004</td>
<td>1526.6</td>
</tr>
<tr>
<td></td>
<td>2006</td>
<td>1477.5</td>
</tr>
</tbody>
</table>

Note: The satellite scenes correspond to spatially different areas, which is why the surface of the classes is.

4. Conclusions

The use of the S-SEBI model permitted estimation of \( E_{T_d} \) on a regional scale in the dry land of Chile’s central-southern zone, which does not have adequate instrumentation to monitor water and energy flows. The model was fed with ASTER images with a high spatial resolution, which was complemented with information from the MODIS sensor and meteorological data in situ of daily maximum and minimum temperatures.

The proposed methodology enabled comparison of the differences between the land use maps and soil types that compose the landscape of the zone, contributing important information regarding water consumption in the territory. Thus, the highest \( E_{T_d} \) rates were obtained in the forest covers with values of 7.3 ± 0.8 and 8.4 ± 0.8 mm d\(^{-1}\) for 2004 and 2006, respectively. The plantations of exotic species had slightly higher rates than the native forests, and their differences were statistically significant. As far as the physical properties of the soil are concerned, it was established that the textural classes and depth are associated with \( E_{T_d} \) following the theoretical patterns associated with water availability in the soil for the plants. It should be pointed out that the time distribution of the pluviometric events and the type of soil dependent on the vegetation cover developed in it were factors of interest to explain the spatial patterns of the \( E_{T_d} \).

These results improve understanding of the spatial patterns of the landscape based on quantitative results regarding the relationship that exists with the water consumption evaluated as ET and the biophysical characteristics of the study zone. This work sets a precedent in the generation of tools to optimize decision-making in the management and planning of water resources in territories of dry land with insufficient measuring instruments, supporting future integrated strategies for dealing with water and soil use. It is worth emphasizing that the results become increasingly relevant because of this being a zone declared as susceptible to climate change, and there are currently problems of water availability, which makes the search for solutions for future adverse scenarios all the more acute.

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